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Combining methods for detection of bycatch hotspot areas of marine megafauna species in and around critical rookeries and foraging areas

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Combining methods for detection of bycatch hotspot areas of marine megafauna species in and around critical rookeries and foraging areas



The rationale



- Conduct a spatial analysis to delineate the overlap and intensity of interaction between animal habitat use and fishing activity through a vulnerability assessment
- Identify bycatch hotspots and implement proper mitigation measures

Bycatch impact



Mortality (entanglement and drowning)

Injuries (sustained from fishing gear)

Population decline

Stress (bycatch survivors may have injuries and experience stress that influences their reproduction and survival probability)

SEABIRDS

Mortality (Direct death due to entanglement)

Population Decline

Disruption of Ecosystem's function (Seabirds are an important indicator of the ecosystem's health and bycatch can affect their population and therefore the entire ecosystem's functioning)

(C) MITIGATION ACTIONS

Technical solutions (bycatch reduction devices, green LED lights on the nets, weighted lines)

Training of fishers on saferelease

Incentives for fishers that avoid marine protected areas

Regulation of high bycatch fishing gear and methods

Bycatch monitoring

Policy/regulations to enforce the use of sustainable fishing practices

Combining methods for detection of bycatch hotspot areas of marine megafauna species in and around critical rookeries and foraging areas

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7 Abstract

6

Bycatch, the incidental catch of non-target species, threatens marine 8 megafauna such as sea turtles and sea birds in the Mediterranean region. 9 Identifying bycatch hotspots is essential to guide mitigation measures and 10 target audiences. In the Mediterranean Sea, South Ionian Sea in Greece 11 is a major marine habitat, including critical nesting areas, for sea turtles, 12 and an important breeding and foraging habitat for sea birds. This work 13 combined methodologies to identify bycatch hotspots through a vulnerabil-14 ity assessment and questionnaire surveys utilising both scientific data and 15 local ecological knowledge (LEK). The study determined the major bycatch 16 hotspots for sea turtles and sea birds, evaluated the potential impact on both 17 species, and discussed mitigation measures to reduce the impact of bycatch 18 and effectively protect this economically and ecologically important ecosys-19 tem. Our approach and outcomes may well contribute to a science-based and 20 LEK included, adaptive management framework regarding the establishment 21 or revision of Marine Protected Areas in the study area and elsewhere across 22

- ²³ critical marine habitats for sea turtles and sea birds.
- 24 Keywords: Bycatch, Sea turtles, Sea birds, Satellite data, mitigation
- ²⁵ strategies, Vulnerability assessment

26 **1. Introduction**

One of the major anthropogenic threats affecting marine megafauna con-27 servation is bycatch (i.e. unintentional catch due to excessive fishing ac-28 tivities) (de la Hoz Schilling et al., 2023; Putman et al., 2020; Domingo 29 et al., 2025). Marine megafauna species encounter various types of fisheries 30 due to their extensive geographic distribution and the use of vast marine 31 areas across different regions (Wallace et al., 2010). The likelihood of in-32 teractions between fisheries and marine megafauna depends on the spatial 33 and temporal overlap between essential habitats for these species and fish-34 ing operations, with the latter involving various fishing techniques and gear 35 types. Despite considerable efforts aiming to address the vulnerability of 36 megafauna to bycatch and identify critical areas where species are prone to 37 bycatch (Lewison et al., 2014; Cardona et al., 2025), scarcity of information 38 for many regions worldwide remains a challenge hindering both sustainable 39 fisheries management and effective conservation of marine megafauna (Cook 40 and Heath, 2018; Fuentes et al., 2023). In the Mediterranean Sea, bycatch 41 is known to affect several marine megafauna species, with sea turtles and 42 sea birds being amongst the most impacted (Izquierdo-Serrano et al., 2022; 43 Karris et al., 2018; Simantiris et al., 2024). 44

Sea turtles are an emblematic group of reptiles broadly distributed across 45 all oceans except polar regions, presenting great importance to marine ecosys-46 tems and the food chain (Hannan et al., 2007; Simantiris, 2024, 2025). Due 47 to their highly mobile nature, sea turtles exploit multiple coastal, neritic, 48 and oceanic habitats at different life stages across their long and complex 49 life cycle (Casale et al., 2018). Yet as highly migratory species, they tra-50 verse between foraging and nesting areas which may be located thousands of 51 kilometers apart (Dujon et al., 2018; Hays et al., 2006; Stokes et al., 2015). 52 Therefore, sea turtles are prone to the bycatch threat deriving from many 53 different fisheries (Casale, 2011; Lewison et al., 2014). In the Mediterranean 54 region, bycatch appears to increase the overall mortality of sea turtle pop-55 ulations (Dimitriadis et al., 2022b; Papazekou et al., 2024b; Agabiti et al., 56 2024). 57

Seabirds constitute a diverse group of more than 400 species, spending 58 part or all of their lives interacting with oceans, e.g., by foraging and migrat-59 ing over them (Harrison et al., 2021). They also constitute one of the most 60 threatened group of birds facing various ecological challenges (Croxall et al., 61 2012; BirdLife, 2018). These marine top predators are nowadays recognized 62 as critically important bioindicators of marine ecosystems that are useful in 63 assessing the environmental disruption and the impacts of climate change 64 on marine biota (Parsons et al., 2008; Mesquita et al., 2015). Indirect mor-65 tality from bycatch in fishing gear is presently a major threat to seabirds, 66 with longline and gillnet fishing being the most impactful fishing practices 67 (Cortés et al., 2017; Dias et al., 2019; Courbin et al., 2024). Despite these 68 considerations, significant gaps of knowledge for the spatial distribution and 69 rate of seabirds' bycatch persist in the Mediterranean, including key areas 70 such as the central and Eastern Mediterranean (Ram'ırez et al., 2024). 71

Identifying the strength and extent of the spatial overlap between high 72 use areas by marine megafauna and human induced threats is often a chal-73 lenging task (Dimitriadis et al., 2022b; Ferreira et al., 2023). The threat 74 of bycatch to sea turtles and seabirds has been assessed using direct, indi-75 rect, and combined approaches aiming to identify mortality hotspots and 76 areas with a high likelihood of species interactions with fisheries. Direct ap-77 proaches to assess the bycatch and mortality rate involve data collection by 78 observers onboard vessels from industrial fishing fleets (e.g. (Cambie et al., 79 2013; Cardona et al., 2025)), reporting by fishers on logbook programmes 80 and questionnaire surveys (e.g. (Tagliolatto et al., 2020; Baldi et al., 2022; 81 Tubbs and Berggren, 2024)). Alternatively, indirect methods for identifying 82 bycatch hotspots and high-risk areas for marine megafauna due to fisheries 83 focus on analyzing the overlap between areas frequently used by the animals 84 and the distribution and intensity of fishing effort (e.g. (Pikesley et al., 2018; 85 Almpanidou et al., 2018, 2021; Hatch et al., 2023; Saüt et al., 2024)). 86

This work, as other studies described above, sets a methodological frame-87 work that combines direct and indirect cost-effective approaches, incorporat-88 ing both scientific data and local ecological knowledge, to identify and evalu-89 ate bycatch hotspots in a case study across a critical area for sea turtles and 90 sea birds at the Mediterranean level, the South Ionian Sea (Greece, Mediter-91 ranean) (Issaris et al., 2012; Karris et al., 2018; Simantiris et al., 2024). In 97 this study, the authors combine of satellite data and questionnaire surveys, 93 and by the use of a vulnerability assessment approach, we aim to spatially 94 delineate habitat use by sea turtles and sea birds, identify the main fishing 95

⁹⁶ grounds and their overlap with habitat use by the animals, and ultimately ⁹⁷ point out bycatch hotspots. Consequently, this study provides valuable infor-⁹⁸ mation on where conservation efforts should be allocated and prioritized and ⁹⁹ suggests a suite of mitigation measures and best practices for the alleviation ¹⁰⁰ of bycatch impacts on marine megafauna.

101 **2. Materials & Methods**

102 2.1. Study area

Zakynthos Island (Ionian Sea, Greece) is home to the second largest 103 nesting population of loggerhead (Caretta caretta) sea turtles in the entire 104 Mediterranean region (Casale et al., 2018), with about 300 females and 100 105 males breading and laying an average of 1200 nests annually (Schofield et al., 106 2017; Margaritoulis et al., 2022). On top of that, nearby marine areas off-107 Zakynthos Island are systematically used by young and adult turtles (Caretta 108 caretta, Chelonia mydas) year-round as foraging areas and migration corri-109 dors (Dimitriadis et al., 2022a; Papazekou et al., 2024a). At the same time, 110 the biggest rookery at the Mediterranean level (Kyparissia Bay) and sev-111 eral other stable nesting sites (around the coastline of Kefalonia Island) are 112 located a few tens of kilometers away from the study area (Casale et al., 113 2018; Dimitriadis et al., 2022b). The study area includes 3 protected areas 114 of the EU habitat and birds Directives (92/43/EEC and 2009/147/EC, re-115 spectively) under the codes GR2210001, GR2210002, GR2210004, as well as 116 the National Marine Park of Zakynthos (NMPZ) established by a Presiden-117 tial Decree in 1999 (Simantiris et al., 2024). NMPZ includes the Strofades 118 island group which is nowadays considered a European breeding population 119 stronghold of Scopoli's shearwater (Calonectris diomedea) and the species' 120 largest colony at a national level since it hosts ca. 5,550 pairs while areas 121 around and off Zakynthos island constitute important foraging areas (Karris 122 et al., 2017; Keller et al., 2020). Moreover, the region is a marine habitat 123 of elasmobranchs, cetaceans, and monk seals (Frantzis et al., 2019; Giovos 124 et al., 2021; Papazekou et al., 2024a; Panou et al., 2023). The Natural Envi-125 ronment & Climate Change Agency (NECCA), is the dedicated unit of the 126 Greek government for the management of these marine and coastal areas. 127

128 2.2. Innovation of previous methodologies

Existing methodologies for identifying fishing fields include analyzing VMS (Vessel Monitoring System) data, Automatic Identification System

(AIS) data, self-reporting, questionnaires, and onboard observer data (Yan 131 et al., 2022; Wang et al., 2024; Mesquita et al., 2024; Maina et al., 2016; 132 Hu et al., 2016; Ram'ırez et al., 2024; Precoda and Orphanides, 2024; Maina 133 et al., 2018; Tzanatos et al., 2005; Moutopoulos et al., 2020). However, rely-134 ing solely on VMS and AIS data has several limitations as to the accessibility 135 and reliability of the data, and the fact that VMS is not required for all fish-136 ing vessels and fisheries, and especially not for small-scale fishing activities 137 (vessels \leq 12 m) (Wang et al., 2024; Thoya et al., 2021; Birchenough et al., 138 2021). On the other hand, the onboard observer and self-reporting method-139 ologies are often determined to be biased and/or incomplete (Gilman et al., 140 2019; Tubbs and Berggren, 2024). Questionnaires have proven significant for 141 estimating the fishing grounds, although the researchers should evaluate the 142 responses to discriminate true from biased responses (Karris et al., 2013). 143 The innovation of this work is the combination of satellite data (proven to 144 be extremely reliable in detecting vessel occurrence (Santamaria et al., 2017; 145 Paolo et al., 2024)), questionnaires for the fishing community of Zakynthos, 146 and existing telemetry data for species occurrence in the identified fishing 147 fields. 148

149 2.3. Questionnaires

A total of 32 small-scale artisanal fishers (SSF), representing more than 150 90% of the overall registered fishers (N=35 with a fishing fleet of 35 vessels) of 151 the local sector (their fishing area extends all around Zakynthos Island up to 152 3 nautical miles from the coast) (Bennett et al., 2020), responded positively 153 to answering a number of predefined questions, from different fishing shelters 154 and harbors around the island (Fig.1). The fishers had an average of 32 years 155 of experience fishing in the area, with their active years varying between 6 156 and 52. 157

158 2.4. Satellite data

SSF usually consists of relatively small vessels (less than 12 meters in total length) that usually operate within the first three nautical miles from the coast and within a restricted range from their home harbor (Calò et al., 2022), without the obligation to report fishing operations through VMS and AIS systems (Regulation (EU) No 508/2014). After collecting spatial information on the fishing grounds from local fishers (questionnaire surveys), and integrating relevant information from scientific literature (Dimitriadis et al.,



Longitude

Figure 1: The harbor locations and number of fishers that responded to questionnaires in Zakynthos Island, Greece

2022b; Karris et al., 2013), which however refers to a lower spatial resolu-166 tion than the one needed herein, we employed a complementary approach 167 for the detection of small scale fishing boats at finest spatial scale by using 168 satellite Synthetic Aperture Radar (SAR) images (Mahdavifard et al., 2022; 169

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Latitude

Satya et al., 2023; Ouchi, 2016). The pros of using satellite data to monitor maritime activity are the ability to cover vast areas, the process is successful regardless of cloud cover conditions, and most importantly, the vessels do not need to cooperate. The process relies on the fact that radar signal is differently reflected by the sea and the vessels, and hence the satellite sensor will receive different signals (Chaturvedi, 2019; Bioresita et al., 2018). Therefore, detecting vessels in the sea becomes possible.

In this study, SAR data from the sensors onboard the Sentinel-1 satellites 177 were used to detect fishing vessels in the area near Zakynthos Island, Greece. 178 The data were used from the Copernicus Collection data on the Google Earth 179 Engine (GEE) following an approach suggested by previous studies as well 180 (Rodr'iguez-Benito et al., 2021; Gascoin, 2019). The approach is based on 181 the creation of a map composite in GEE from the Sentinel-1 image collection 182 provided by Copernicus (Gorelick et al., 2017). After carefully selecting the 183 area of interest, a few modifications were made to the original script provided 184 by Gascoin (2019) in order to acquire the data on the occurrence of ships in 185 the study area (Fig.2). The GEE script was used to detect ship occurrence 186 within the study area for a time-series between 2014 and 2024. Due to the 187 difference in the scattering signal of sea and vessels, the script provided a 188 map of the study area, created from the Sentinel-1 images (Sentinel-1 Image 189 Collection: COPERNICUS/S1-GRD), with black pixels representing the sea 190 surface and white pixels representing vessels (Fig.2). The data were filtered 191 to get images collected in interferometric wide (IW) swath mode and VH 192 (transmitter-receiver) polarization at both ascending and descending orbit 193 passes. Finally, the images were filtered to acquire images from the same 194 angles of view and the generated map was exported as GeoTIFF to be further 195 analyzed in Matlab. 196

In Matlab, the GeoTIFF images were imported and the data points with values greater than zero were determined. The values equal to zero were assumed to correspond to the reflectance of the sea surface. The land was masked to avoid adding bias to our results. To assess the size of vessels, all the connected points were assigned a polygon to estimate their size. To identify the small-scale vessels and discriminate them from other types of vessels (e.g. cargo) we used the size of the vessels (between 2-12 m in length).

204 2.5. Telemetry data for species

Data on the occurrence of sea turtle species within the study area were extracted from the OBIS-SEAMAP database, the largest data center for sea



Figure 2: The analysis used in GEE to geographically define and extract the vessels' data

turtle occurrence and distributions in the world (Halpin et al., 2009). As 207 expected, the greatest number of occurrences is noticed in the Laganas Bay 208 in Zakynthos where several thousands of sea turtles are gathered each year 209 during the nesting period (Simantiris et al., 2024; Simantiris, 2024) (Fig.3). 210 In accordance with previous studies (Schofield et al., 2013a), several presences 211 are seen in the surrounding areas proving that sea turtles are present on the 212 island year-round. In addition, data on the use of the pelagic and coastal 213 zones of the study area were also used for the Scopoli's Shearwater breeders in 214

the Strofades colony. The spatial data were based on the use of waterproofed 215 GPS data loggers storing tracking information on 30 different breeders. The 216 loggers were attached to the four central tail feathers using TESA tape and 217 configured to record positions every 15 min. Weighing a total of 20-23 g, 218 the loggers (45 imes 32 imes 18 mm) comprised slightly more than 3% of the 219 mean body mass, which constitutes the recommended threshold for ensuring 220 the elimination of any possible effect on their movement behavior (Phillips 221 et al., 2004; Passos et al., 2010). Data collection was implemented during 222 different breeding seasons between 2009 and 2018. The tracked birds were 223 removed from their breeding burrows between mid of July and early August 224 when the majority of the chicks had hatched and were about 1–2 weeks old. 225 On returning to their nests over the following days, the birds fitted with 226 GPS loggers were recaptured after food provision to chicks; the loggers were 227 removed and data were downloaded and stored. Generally, some of the main 228 foraging grounds according to GPS data are located in coastal areas around 229 Zakynthos Island and off the north-western Peloponnese (Karris, 2014; Karris 230 et al., 2018) (Fig.4). The high density of GPS locations found in the vicinity 231 (within 2-3 nm) of the Strofades colony is not considered an indication of a 232 core foraging area but as bird aggregations just before visiting their nesting 233 sites to feed the chicks during the night. It is known that Procellariiform 234 seabird species such as Scopoli's Shearwater exhibit nocturnal behaviour as 235 an adaptation strategy to avoid terrestrial predators. As a response to that 236 threat, they tend to form flocks or "rafts" during dusk, just before coming 237 ashore to their nesting sites at night (Karris et al., 2018; Rubolini et al., 238 2015). 239

240 2.6. Vulnerability assessment

A vulnerability assessment was carried out in an effort to identify the 241 bycatch threat level for sea turtles and sea birds within the study area. In 242 order to evaluate the vulnerability of each location to bycatch, the method-243 ological approach described by Cuevas et al. (2019) was followed. This ap-244 proach calculated the ecological vulnerability of selected species based on 245 quantified sensitivity data (degree of impact from specific threat), expected 246 threat (occurrence of a specific threat), and a stability factor (environmental 247 and/or anthropogenic features that influence the effect of the threat). The 248 pixel dimensions used that represent the spatial resolution of the dataset 249 were 50x50 m. The following equations were used for the estimation of the 250 cumulative vulnerability: 251



Figure 3: The occurrence of sea turtles in the study area from satellite tag data

$$V = Sens * E_{Th} - SC \tag{1}$$

where V is the vulnerability for each species, *Sens* is the sensitivity to a specific threat, E_{Th} , is the expected threat, and *SC* is the stability factor. The expected threat was given a value between 0 and 1 based on the fishing effort from the satellite data (1 being a high occurrence of fishing vessels and 0 a low occurrence derived from a density spatial interpolation based on the satellite telemetry data for vessel occurrence), and the stability factor was 0 for regions outside marine protected areas (MPAs) and 1 for MPAs.



Figure 4: The occurrence of Scopoli's Shearwater breeders originating from the Strofades colony (seen with a star symbol) during their foraging distribution in the study area according to GPS data (sampling period: 2009-2018).

$$Sens = \sum_{i=1}^{N} (\lambda_i * Att_i)$$
(2)

where λ_i is the weight of each attribute Att_i. The attributes here represent

the intensity of use of different locations by each species (i.e., nesting or

feeding grounds). The weights were given a value of 0.3 for nesting and

0.6 for feeding grounds according to Cuevas et al. (2019). Identification of

feeding and nesting grounds for both species was based on previous research

effort in the study area (Almpanidou et al., 2022; Karris et al., 2018). The

attributes were given a value between 0 and 1 based on the presence of the

species (1 being a high occurrence of individuals and 0 being a low occurrence

derived from a density spatial interpolation based on the satellite telemetry

 $CV = \sum_{i=1}^{N} (V_i)$

(3)

where *CV* is the cumulative vulnerability for both species. The CV was rescaled to include values between 0 and 1.

271 3. Results

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272 3.1. Identifying fishing grounds

data) from the satellite telemetry data.

According to the local fishing community of Zakynthos, 68% of the fishers reported fishing during the night, 26% during the day and night, and only 6% reported fishing during the day (Fig.8). The fishing fields for small-scale fishers include the regions of Porto Vromi, Mizithres, Laganas Bay, Tsilivi, and the channel between the islands of Zakynthos and Kefalonia.

The analysis of the GeoTIFF images showed several thousand vessel lo-278 cations within the study area for the selected period (Fig.2). The analysis 279 showed that most vessels were smaller than 50 m in length, with a small 280 number exceeding 100 m (ferry/cargo ships). The total number of vessel 281 occurrences between 2-12 m for the timeframe between 2014-2024 was ap-282 proximately 11,000 (Figs.6,2). The projection of their location in the map 283 revealed two regions of small-scale vessel aggregations, one on the NE coast 284 of Zakynthos and one between the northernmost location of Zakynthos and 285 the S Kefalonia Island (Fig.5). These locations are therefore identified as the 286 hotspots of small-scale fishing activity in the study area in accordance with 287 the information provided by local fishers but in a finer spatial detail. 288

289 3.2. Identifying bycatch hotspots

290 Considering bycatch, the questionnaires showed that more than half of 291 the fishers have caught at least one turtle in their fishing gear, with 19% of



Figure 5: The identified vessel occurrence between 2014-2024 in the study area from the analysis of Sentinel-1 SAR images, Dar-grey background shows the region of known fishing activities with lighter-grey areas showing less active regions.

the fishers reporting to have caught several sea turtles during their active fishing years. A smaller amount of fishers reported 2-3 sea turtles (10 and 10% respectively). According to the fishers, bycatch usually takes place in the



Figure 6: The length of identified vessels in the study area between 2014-2024

summer months, and between May and October (Fig.7). The areas of high 295 bycatch as reported by the local fishermen are shown in Fig.8. According to 296 the analysis of Sentinel-1 SAR images, the occurrence of vessels in the study 297 area occurs in aggregations in two locations (Fig.5). The first location is 298 near Zakynthos on the northern side of the island near the Alykanas harbor. 299 The second and larger one is located between Zakynthos and Kefalonia in 300 the Lourdas Bay. Considering the overlap of the presence of sea turtles and 301 sea birds in these regions, as identified by telemetry data, and the fishing 302 grounds, as identified by the satellite data and the questionnaires, the two 303 areas can be considered significant hotspots for bycatch of the aforementioned 304 marine species (Figs.3,4). 305

306 3.3. Vulnerability map

Based on equations:1,2,3, the total cumulative vulnerability was assessed for the impact of bycatch on sea turtles and sea birds in the study area



Figure 7: Answers of fishers in main questions involving bycatch in Zakynthos Island, Greece

(Fig.9). The most vulnerable region was identified to be in the area between
the islands of Zakynthos and Kefalonia, an area that is experiencing high
fishing activity and the occurrence of both sea turtles and sea birds, with
some other areas in the coasts of Zakynthos island exhibiting slightly higher
values than the other regions.

314 **4. Discussion**

Small-scale fisheries in the Mediterranean comprise more than 80% of the fishing vessels, 50% of employment onboard vessels, and around 30% of revenues (FAO, 2020), while it is crucial for the welfare of the coastal population of developing countries (Teh and Sumaila, 2013). According to Fao et al. (2018), in the last 7 decades, the global consumption of fisheries products has increased exponentially and is expected to keep increasing by 1.5%



Figure 8: Bycatch hotspots according to the local ecological knowledge of the fishers in the study area

annually. In the Mediterranean region, most fishing activities occur by small 321 vessels using various types of fishing gear (Lleonart and Maynou, 2003). Due 322 to the intensive needs of the population of the Mediterranean region for the 323 consumption of seafood products, the fishing industry is expanding leading 324 to the overexploitation of resources and the degradation of the marine envi-325 ronment (Colloca et al., 2017; Lotze et al., 2011). Due to the extensive fishing 326 activities, marine organisms such as sea turtles, elasmobranchs, cetaceans, 327 and sea birds are impacted due to bycatch (Virgili et al., 2024). Bycatch is 328 caused by the overlapping of fishing grounds with the species' marine habi-329

Journal Pre-proof



Figure 9: The spatial representation of the total cumulative vulnerability assessment for bycatch for sea turtles, sea birds, and the total cumulative vulnerability for both species in the study area

tats, leading to high mortality rates, especially for sea turtles (Caretta caretta, 330 Chelonia mydas), monk seals (Monachus monachus), whales (Ziphius cavi-331 rostris, Physeter macrocephalus, Megaptera novaeangliae), dolphins (Stenella 332 coeruleoalba, Tursiops truncatus, Delphinus delphis) and seabirds (Calonec-333 tris diomedea)(Papazekou et al., 2024a; Li Veli et al., 2024; Virgili et al., 334 2024; Tomás et al., 2008; Karris et al., 2018). Bycatch can account for up to 335 40% of the fishing activity's product, with a smaller percentage comprising of 336 megafauna species (McCauley et al., 2015; Allman et al., 2021; Wallace et al., 337 2011; Dulvy et al., 2021). Hence, bycatch mitigation measures are essential 338 to ensure the conservation and sustainability of the marine environment and 339

³⁴⁰ ecosystem of our planet.

341 4.1. Impact on Sea turtles

According to the telemetry data, sea turtles are present in every region 342 around the island, with higher densities reported within Laganas Bay due 343 to the nesting activity (Fig.3). The areas around the island and between 344 the islands of Kefalonia and Zakynthos are considered mating and foraging 345 areas (Casale et al., 2018; Papazekou et al., 2024b,a) in close proximity to the 346 nesting beaches of Mounda Bay, at the southeast part of Kefalonia island. Sea 347 turtles are most commonly reported between early spring when mating starts 348 (Schofield et al., 2013b), and till the end of October. As commented by the 349 fishers (Figs.7,8) and verified by the satellite data (Fig.5), the fishing grounds 350 with higher intensity include the regions between the northern Zakynthos and 351 the southern Kefalonia islands. The questionnaires also showed that in these 352 regions, sea turtle bycatch is frequent, especially at night, between May to 353 October. Considering the increasing trend of fishing activities (Fao et al., 354 2018), the increasing nesting activity on Zakynthos island (Margaritoulis 355 et al., 2022), and the movement of Mediterranean sea turtles from the eastern 356 basin towards the central and western due to the impact of climate change 357 (Simantiris, 2024), the identified hotspots for bycatch in Zakynthos island 358 may pose a significant threat for the species as bycatch will also increase 359 accordingly, leading to higher numbers of stranded sea turtles and other 360 species in the region (Papazekou et al., 2024a). 361

362 4.2. Impact on Seabirds (Scopoli's Shearwater breeders)

Similarly to this study, Karris et al. (2013) surveyed Zakynthos, Greece, to 363 identify bycatch rates in the southern Ionian Sea. The authors distributed a 364 questionnaire to the majority of small-scale fishers in the harbors of the island 365 and reported significant incidental catches of Scopoli's Shearwater and in a 366 lesser extent of Mediterranean Shag due to commercial longline and gillnet 367 fishery gears. Also, the fishers provided direct information on the fishing 368 grounds where incidental catches of seabirds mainly occurred (e.g., coastal 369 regions of Zakynthos Island and southern coastal area of Kefalonia Island) 370 that matched the findings of the current study. Moreover, temporal analysis 371 of the incidental bird mortality showed that seabirds were more susceptible to 372 being trapped in fishery gears set around sunrise during spring and summer. 373 According to Karris et al. (2013) the estimated annual incidental mortality of 374 Mediterranean Shags in bottom longlines and nets represents approximately 375

3.0-5.1% of the pairs breeding in Southern and Central Ionian Sea (HOS 376 unpublished data). Similarly, 495 Scopoli's shearwaters were estimated to 377 be caught in longlines which represents 1.7-2.0% of the local population. 378 Although bycatch of the shearwaters during the pre-breeding period in early 379 May could affect birds that migrate via the Southern Ionian Sea, the highest 380 bycatch rates occur during summer months, when it can be assumed that 381 the caught birds mainly originated from the Strofades colony. Consequently, 382 bycatch mortality of Scopoli's shearwater could be considered a potential 383 risk for the local colony by taking into consideration that this marine top 384 predator shows long-term mate fidelity as well as biparental care during the 385 incubation of the single egg per nest and the chick-rearing duties. 386

387 4.3. Mitigation measures

Bycatch is a major threat to both the conservation and sustainabil-388 ity of marine life, and the fishing gear used by fishers around the globe 389 (Agyekumhene et al., 2014; Gautama et al., 2022; Cardona et al., 2025). 390 Hence, several approaches have been used in an attempt to reduce the im-391 pact of bycatch on marine megafauna. The best practices to mitigate bycatch 392 depend on the geographical area, fishing gear types, bycaught species, impor-393 tance for the local population, existing legislation, and regional management 394 authorities (Squires et al., 2021). Especially in the case of small-scale fish-395 eries, as in Zakynthos Island, mitigating bycatch can be very challenging. 396

Existing approaches involve the following methods: i) the combination of 397 bycatch reduction devices (BRD) on board fishing vessels with the train-398 ing/education of fishers on good practices, technical solutions, and eco-399 labeled to achieve bycatch mitigation (Virgili et al., 2024), ii) the use of green 400 LED lights on the nets to reduce the bycatch of sea turtles and weighted lines 401 to reduce the bycatch of sea birds (Gautama et al., 2022; Løkkeborg, 2011), 402 iii) awards as incentives for fishers that reduce the bycatch of marine mam-403 mals and avoid marine protected areas (Lent and Squires, 2017; Macedo 404 et al., 2019), iv) the closure of specific areas (Squires et al., 2018), v) the 405 ban of specific fishing gear (Sala, 2016), vi) the alterations of existing fishing 406 gear and methods (Senko et al., 2017; Squires et al., 2018; Fitzgerald, 2013; 407 Atkins et al., 2013; Virgili et al., 2018; Lyle and Tracey, 2016; Henry et al., 408 2024), vii) the use of technological approaches with innovative devices for 409 monitoring bycatch (Wakefield et al., 2018; Bartholomew et al., 2018) and 410 alienating specific species from the fishing gear (Duarte et al., 2019; Jefferson 411 and Curry, 1996; Wang et al., 2010), viii) the implementation of awareness 412

and training campaigns targeting local fishing communities (Squires et al., 413 2018; Senko et al., 2017; Bretos et al., 2017), ix) the introduction of eco-414 labeling (Selden et al., 2016; Lent and Squires, 2017; Bellchambers et al., 415 2014; Christian et al., 2013; Berninsone et al., 2018), x) the use of dynamic 416 ocean management methods (DOMs) where fishers, NGOs, authorities, and 417 managers collaborate to evaluate the movement and distribution of pelagic 418 species to adjust the spatiotemporal fishing grounds (Dunn et al., 2016; Lewi-419 son et al., 2015; Siders et al., 2024), x) and the implementation of observer 420 programs (Bellchambers et al., 2014; Lent and Squires, 2017), among others. 421 In Zakynthos, the most common approach to avoid bycatch is the onboard 422 release of bycaught organisms such as sea turtles and sea birds, a common 423 practice that is taking place due to the training of the fishers by local NGOs 424 and the fishers' awareness. According to the fishers' responses (Fig.7), more 425 than 60% have caught a sea turtle in their fishing gear and have released it. 426 They disclosed that if the sea turtle was alive, the release would take place 427 even if it meant causing damage to their gear, while if the sea turtle was al-478 ready dead, they would release it later in order to cause as less as possible to 429 their gear. On board release is a common practice that, although voluntary, 430 is highly significant for the conservation of marine organisms, but differen-431 tiates between species (Wosnick et al., 2023). Nevertheless, combined with 432 education and workshops, best practices for the release of bycatch products 433 can be communicated to the majority of fishers around the globe (Wosnick 434 et al., 2023). The authors suggest that studies involving the use of streamer 435 (tory) lines in longline vessels to evaluate the protective effect of this setup 436 on seabirds are critical for finding the effectiveness of this bycatch mitigation 437 measure for seabirds. This will allow evaluating its effectiveness towards the 438 reduction of the loss of seabirds in longline fishery following other relevant 439 studies (e.g. (Cortes and Gonzalez-Solis, 2018)). Moreover, the need to in-440 clude more marine megafauna species data in the vulnerability assessment for 441 bycatch is important to define specific bycatch hotspots and assist in inform-447 ing conservation plans. The information presented here supports the need for 443 conservation, education, and engagement actions in the region of the Ionian 444 Islands for the preservation of the marine environment and the mitigation of 445 bycatch, especially considering the high ecological and economic importance 446 of the region due to its biodiversity and the role in fisheries. 447

448 **5.** Conclusions

In the Mediterranean region, bycatch is a major threat to marine life. In 449 Zakynthos Island, Greece, bycatch is known to have a significant impact on 450 sea turtles and seabirds, among other species. This work combined satellite 451 data, questionnaires, and GPS data to identify the fishing fields and evaluate 452 the interaction with sea turtles and seabirds. The current work reports one 453 important marine area with systematic fishing activities, which is also a 454 marine habitat for sea turtles and sea birds, and verified through the local 455 fishing community and a vulnerability assessment as a bycatch hotspot. It 456 also provides a useful methodological tool for researchers using different data 457 sources to identify bycatch hot spot areas of marine protected species that are 458 susceptible to incidental mortality on fishery gears. At a national level, the 459 findings of the current study will also serve the need to advise conservation 460 planning in MPA designation in the study area and elsewhere. 461

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The author declares no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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469 8. CRediT author statement

Nikolaos Simantiris: Conceptualization, Methodology, Software, Re-470 sources, Investigation, Data Curation, Writing-Original Draft, Visualization, 471 Formal analysis, Project administration. Charalampos Dimitriadis: Data 472 Curation, Writing, Review & editing. Stavros Xirouchakis: GPS data 473 collection for seabirds, Writing-Original Draft. Marios-Dimitrios Voul-474 garis: GPS data collection for seabirds, Writing-Original Draft. Evangelia 475 Beka: Resources, Write, Review & editing. Martha Z. Vardaki: Writ-476 ing, Review & editing Georgios Karris: GPS data collection for seabirds, 477 Writing-Original Draft, Supervision. 478

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- Combining satellite data and questionnaires assist in determining fishing grounds
- Bycatch hotspots for sea turtles and sea birds in Zakynthos, Greece were determined
- There is a need for conservation, education, and engagement actions
- This work contributes to the establishment of a new national marine park

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Declaration of interests

☑ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

□ The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

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